

Ecology of hawksbill turtles *Eretmochelys imbricata* on a western Caribbean foraging ground

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Keywords: marine turtle, hawksbill, *Eretmochelys imbricata*, tagging, habitat use, behavior

This manuscript has been provisionally accepted for publication in
Chelonian Conservation and Biology.

Abstract

1 We present results of an inwater research program focusing on basic ecology of
2 juvenile hawksbill turtles *Eretmochelys imbricata* in the Cayman Islands. We made
3 206 captures of 135 hawksbills in Little Cayman (LC) and 103 captures of 97
4 hawksbills in Grand Cayman (GC). The Cayman Islands aggregation demonstrated a
5 broad size distribution (20.5–62.6 cm straight carapace length), slow growth rate (3.0
6 ± 0.9 cm yr⁻¹), and multiple recaptures, suggesting long-term residence in some
7 individuals. Demonstrated home range was small (mean distance from capture to
8 recapture 545 ± 514 m, range 2–2080 m) although an international tag return
9 suggested a long-range developmental migration. Vertical features provided important
10 habitat in LC, and larger turtles were generally captured in deeper waters. Behavior at
11 sighting varied by habitat: resting, swimming, and feeding were observed in coral
12 reef, reef wall, and hardbottom colonized by sponges and gorgonians, and resting was
13 frequently observed in uncolonized hardbottom. Images obtained from underwater
14 photographers enhanced understanding of hawksbill foraging behavior: turtles fed on
15 sponges (particularly the leathery barrel sponge, *Geodia neptuni*), by scraping the
16 reef, and occasionally by consuming thimble jellyfish *Linuche unguiculata*. Intra- and
17 interspecific interactions were recorded: an apparently commensal feeding
18 relationship was noted with gray *Pomacanthus arcuatus*, French *Pomacanthus paru*,
19 and queen angelfish *Holacanthus ciliaris* and aggressive, possibly territorial,
20 interactions between hawksbills were observed. We also documented causes of injury
21 and mortality in the study area – including legal, illegal and incidental take, vessel
22 collisions, hurricanes, and natural predation.

Introduction

23 Hawksbill turtles *Eretmochelys imbricata* are a migratory species of conservation
24 concern. Adults travel hundreds or thousands of kilometers from foraging grounds to
25 breeding areas (Horrocks et al. 2001; Troëng et al. 2005; van Dam et al. 2008) and
26 neonates are broadly dispersed by ocean currents (Carr 1987; Musick and Limpus
27 1997). Following the oceanic stage, juveniles recruit to neritic habitats including coral
28 reef, hardbottom, seagrass bed, and cliff-wall (Musick and Limpus 1997). Recent
29 genetic studies suggest that juveniles on foraging aggregations originate from multiple
30 nesting beaches (Bowen et al. 1996; Bass 1999; Díaz-Fernández et al. 1999; Bowen et
31 al. 2007a; Velez-Zuazo et al. 2008) and as hawksbills are commercially-valuable,
32 management of mixed stocks has been the subject of considerable controversy
33 (Bowen et al. 2007b; Godfrey et al. 2007; Mortimer et al. 2007a; Mortimer et al.
34 2007b).

35 While movements across geopolitical boundaries attract considerable attention (e.g.
36 Meylan 1999), juvenile hawksbills appear to remain resident on foraging grounds for
37 extended periods (Limpus 1992; van Dam and Diez 1998a; León and Diez 1999;
38 Sanches and Bellini 1999), where local conditions determine survival. Inwater capture
39 has provided information on size distribution and condition index (León and Diez
40 1999; Diez and van Dam 2002) and growth rate has proven to be habitat dependent
41 and highly variable among study sites, suggesting variation in time to maturity (León
42 and Diez 1999; Diez and van Dam 2002; IUCN 2002). Thus, monitoring demographic
43 parameters from a variety of locations will aid in understanding population dynamics
44 and evaluating resilience to harvesting.

45 Studies of hawksbill habitat use and behavior on foraging grounds may also elucidate
46 ecological roles (León and Bjørndal 2002) and susceptibility to threats. For Caribbean
47 hawksbills, ultrasonic tracking, point of capture habitat assessment, and benthic
48 habitat mapping have begun to illuminate home range and habitat use (van Dam and
49 Diez 1998a; León and Diez 1999; Cuevas et al. 2007). Deployment of time depth
50 recorders (TDRs) has provided data on depth utilization (van Dam and Diez 1996;
51 van Dam and Diez 1997; Blumenthal et al. 2009), yet inwater activities cannot be
52 determined from dive profiles alone (Seminoff et al. 2006). In order to study behavior,
53 increasingly sophisticated technologies – such as video-linked time depth recorders
54 (Heithaus et al. 2002; Seminoff et al. 2006), multi-sensor archival tags (Wilson et al.
55 2008), and inter-mandibular angle sensors (Hochscheid et al. 2005; Houghton et al.
56 2008) – are being developed and applied. However, direct inwater observation
57 (Houghton et al. 2003; Schofield et al. 2006) may represent a complementary and
58 substantially under-utilized method in marine turtle research, considering its potential
59 to offer insights into inwater activities and aid in the interpretation of data gathered
60 through instrumentation.

61 Despite an increasing number of inwater studies, national and international
62 conservation efforts are hindered by a lack of basic demographic and ecological data
63 on immature hawksbills (Mortimer and Donnelly 2007). While juvenile hawksbills
64 are often sighted in the waters surrounding the Cayman Islands (Bell et al. in press),
65 basic ecology and management needs of this aggregation have not previously been
66 assessed. Here, we present results of a seven-year monitoring program, focused on
67 providing relevant biological data from a Caribbean foraging ground. We conducted a
68 capture-mark-recapture study, integrated point of capture habitat assessments with
69 benthic habitat mapping, collected stranding data, and made direct observations of

70 turtle behavior at sighting. Additionally, in order to supplement our observations, we
71 requested photographs from recreational and professional underwater photographers –
72 documenting marine turtle habitat use, diet, and behavior. Thus, through diverse
73 methods, we aimed to elucidate hawksbill management requirements within the
74 foraging ground as well as implications for regional management.

Methods

Study site

75 The Cayman Islands are located in the western Caribbean Sea, approximate 240 km
76 south of Cuba (Fig. 1). For this study, two sampling sites were selected: Bloody Bay,
77 Little Cayman (19.7° N, 81.1° W) and western Grand Cayman (19.3° N, 81.4° W). A
78 narrow shelf surrounding each island consists of coral reef, hardbottom and other
79 habitats. In many locations, a former seacliff (“miniwall”) begins at depths of 8–10 m,
80 marking the transition from shallow to deep terrace reefs. From the edge of the shelf
81 (beginning at depths of 6–20 m), the near-vertical “reef wall” and deep slope extend
82 to abyssal depths (Logan 1994; Roberts 1994). Both the miniwall and the reef wall are
83 characterized by prolific coral reef colonization (Logan 1994; Roberts 1994).

Capture methodology

84 Hawksbill turtles were hand-captured by snorkelers (who swam in teams of two or
85 were towed approximately 10 m behind a small boat). Catch per unit effort was
86 recorded as hawksbills sighted per hour that observers were towed. For each sighting,
87 GPS location, habitat type, water depth, turtle activity, and estimated turtle size were
88 recorded (regardless of whether a turtle was captured). In order to qualitatively
89 supplement observations, we requested photographs of marine turtle habitat use and
90 behavior from underwater photographers.

Tagging

91 To allow individual identification, all captured turtles were tagged according to
92 standard protocols: a metal inconel tag was applied to the posterior edge of each front
93 flipper and a Passive Integrated Transponder (PIT) tag was injected into the shoulder

94 muscle (Balazs 1999). Additionally, to prevent individuals from being captured more
95 than once per capture occasion, a white grease pen was used to apply a temporary
96 mark to the carapace of each turtle.

97 Tag retention for inconel and PIT tags was calculated according to the equation
98 $P_i = b_i / (a_i + b_i)$, where i is the elapsed time in whole years since tag application, p_i is the
99 probability of tag loss i years after attachment, a_i is the number of tags present i years
100 after attachment, and b_i is the number of tags lost i years after attachment (Limpus
101 1992; Bellini 2001). All tags were marked with a return address: the Archie Carr
102 Center for Sea Turtle Research (from 2000 to 2002) and Wider Caribbean Sea Turtle
103 Conservation Network (from 2002 to present). For each recaptured turtle, time at
104 large was noted and straight-line distance from GPS capture site to GPS recapture site
105 was calculated using Hawth's Analysis Tools for ESRI ArcGIS.

Size frequencies, condition index, and growth

106 Measurements of mass and straight carapace length (SCL, measured from the center
107 of the nuchal notch to the tip of the posterior-most marginal scute) were used to
108 determine size distribution and calculate morphometric relationships (van Dam and
109 Diez 1998b), growth rate (Bjorndal and Bolten 1988) and body condition index (mass
110 SCL^{-3}) (Bjorndal et al. 2000). In order to determine the accuracy with which turtle
111 size could be estimated inwater, for captured turtles we compared estimated size (at
112 first inwater sighting) and measured size, calculating mean difference from the
113 absolute values.

Habitat mapping

114 Benthic habitat maps were produced from orthorectified and georectified true color
115 aerial photography (0.12 m resolution), using the NOAA habitat digitizer extension
116 for ESRI ArcGIS v. 9.2. Habitat categories included coral reef (spur and groove,
117 aggregate reef, patch reef, and reef wall), colonized hardbottom (10–70% colonization
118 by sponges and gorgonians), uncolonized hardbottom (<10% colonization), rubble
119 and sand.

Threats

120 We collected data on hawksbill turtles captured in the legal marine turtle fishery (Bell
121 et al. 2006) as well as on injured or dead hawksbills reported to Department of
122 Environment (the agency responsible for responding to marine turtle strandings in the
123 Cayman Islands). Where possible, necropsies were performed to determine cause of
124 death.

Results

125 From 2000–2007, we made 206 captures of 135 individual hawksbill turtles in Little
126 Cayman and 103 captures of 97 hawksbills in Grand Cayman. Species composition of
127 the aggregation was primarily hawksbills: limited sightings and captures were made
128 of juvenile green turtles at both sites, and one juvenile loggerhead was sighted but not
129 captured in Little Cayman. Catch per unit effort (mean \pm SD) in Little Cayman (3.15
130 ± 0.98 hawksbill sightings per hour towing) was significantly greater than Grand
131 Cayman (1.6 ± 0.45 hawksbill sightings per hour towing) (Mann Whitney $u=9.000$,
132 $p=0.0001$).

Size distribution and body condition index

133 Straight carapace length (mean \pm SD) was 33.7 ± 8.6 cm for hawksbills captured in
134 Little Cayman (Fig. 2a, $n=125$ individuals for which a measure of notch-to-tip straight
135 carapace length was obtained) and 31.4 ± 7.4 cm for hawksbills captured in Grand
136 Cayman (Fig. 2b, $n=93$ individuals) and size range for the aggregation was 20.5–62.6
137 cm. Condition index (10^{-4} kg cm⁻³) for Little Cayman (1.25 ± 0.17) and Grand
138 Cayman (1.24 ± 0.18) was not significantly different (Mann Whitney $u=7824$, $p>0.05$,
139 $n=268$). Comparison of estimated with actual size for captured turtles indicated that
140 size could typically be estimated to within 10 cm (mean accuracy \pm SD: 4.66 ± 3.67
141 cm, $n=85$).

Tag retention and recaptures

142 Over the duration of the study, tag retention was nearly 100% for each tag (inconel:
143 98%; PIT 100%), ensuring near certainty that triple-tagged turtles would remain
144 individually identifiable for extended periods. In Little Cayman, ultrasonic tracking

was used to facilitate recovery of hawksbills instrumented with time depth recorders, resulting in recapture of 19 of 21 individuals (Blumenthal et al. 2009). Excluding recaptures facilitated by ultrasonic tracking, in Little Cayman 72% of hawksbills were captured once, 19% were captured twice, 7% were captured three times, 2% were captured four times, and 1% were captured five times. In Grand Cayman, 88% were captured once and 12% were captured twice.

Local recaptures occurred at intervals ranging from 11 days to 7.3 years. Distance traveled (mean \pm SD) from first capture to last capture was 545 ± 514 m, range 2–2080 m, $n=57$ (Fig. 3). There was no correlation between distance traveled and time at large (Spearman's $r=0.06$, $p>0.05$) or distance traveled and turtle size (midpoint between straight carapace length at capture and straight carapace length at recapture) (Spearman's $r=0.09$, $p>0.05$). One hawksbill turtle was tagged in Little Cayman and recaptured 6.7 years later in La Mosquitia, Honduras (likely as an opportunistic capture in an artisanal lobster fishery). Size of this animal at original capture in the Cayman Islands was 46.5 cm straight carapace length and size at recapture was not reported.

Growth and morphometrics

Mean growth rate (of individuals, from first capture to last capture) was 3.0 ± 0.9 cm yr^{-1} ($n=37$ increments of >1 year). Straight carapace length and growth rate were significantly correlated (Spearman's $r=-0.43$, $p<0.01$) (Fig. 4a) and the correlation between straight carapace length and mass was highly significant (Linear regression, $r=0.96$, $p<0.0001$) (Fig. 4b).

Behavior

166 A wide range of behaviors were recorded at sighting. Turtles were observed resting
167 (33%), swimming (31%), breathing (20%), feeding (10%), hovering (3%), fleeing
168 (2%), and fighting (1%) (n=317 observations).

169 We were able to broadly assess how turtle activity varied according to habitat (Fig. 5,
170 n=272 observations for which both habitat and activity were obtained): turtles fed in
171 colonized hardbottom, miniwall, reef, and reef wall, while resting was common in
172 uncolonized habitats. In deeper areas (such as off edge of the wall in Little Cayman),
173 turtles were more likely observed breathing at the surface.

174 In order to qualitatively supplement behavioral observations, more than 500
175 photographs documenting habitat use and behavior were obtained from underwater
176 photographers. During the day, hawksbills were observed in colonized hardbottom
177 (Fig. 6a), reef (Fig. 6b), and reef wall (Fig. 6c) and at night, hawksbills were seen
178 resting on the bottom and wedged under ledges (Fig. 6d). Feeding behaviors included
179 scraping the surface of the reef (Fig. 6e) and eating sponges (primarily the leathery
180 barrel sponge *Geodia neptuni*). Gray *Pomacanthus arcuatus*, French *Pomacanthus*
181 *paru*, and queen *Holacanthus ciliaris* angelfish were documented feeding on sponges
182 in association with hawksbills, nibbling sponges where the interior tissue was exposed
183 (Fig. 6f) or eating crumbs dropped by turtles (Fig. 6g). Additionally, hawksbills were
184 observed feeding on occasional summer swarms of thimble jellyfish *Linuche*
185 *unguiculata* (Fig 6h).

Habitat use

186 When GPS sighting locations were scaled according to turtle size, there was no
187 apparent structuring of horizontal habitat use (Fig. 7a). However, a weak but highly

significant correlation was observed between turtle size and depth of the water at capture (Little Cayman: Spearman's $r=0.35$, $p<0.001$; Grand Cayman Spearman's $r=0.44$, $p<0.001$) (Fig. 7b).

In Little Cayman, turtles were sighted in coral reef (20%), reef wall (26%), colonized hardbottom (39%), and uncolonized hardbottom, rubble, or sand (14%). In Grand Cayman, turtles were sighted in coral reef (57%), colonized hardbottom (17%), and uncolonized hardbottom, rubble or sand (26%). When spatial data were integrated with benthic habitat mapping, sightings and captures of hawksbill turtles were dispersed in Grand Cayman (Fig. 8a) and clustered near the miniwall and reef wall in Little Cayman (Fig. 8b).

Threats

Injury and mortality ($n=41$ documented incidents during the course of the study) resulted from anthropogenic (61%), natural (30%), and unknown (10%) sources. Anthropogenic threats included legal (24%) and illegal take (17%), vessel collision (2%), and incidental capture, including entanglement in fishing line and ingestion of fishhooks (17%). Natural sources included hurricanes (15%) and possible shark inflicted wounds (15%). Additionally, photographic evidence was obtained documenting the presence of hawksbill scutes in the stomach of a tiger shark *Galeocerdo cuvier* captured by fishermen in Grand Cayman in 2002. There is also potential for anthropogenic disturbance: Bloody Bay is heavily utilized for recreational diving and 98% of hawksbill sightings in Bloody Bay occurred within 200 m of a dive mooring.

Discussion

Demographics

209 Size distribution of captured hawksbills suggests that the Cayman Islands provide
210 developmental habitat. Generally, hawksbills are first documented on foraging
211 grounds at 20–35 cm curved carapace length (Musick and Limpus 1997; van Dam and
212 Diez 1998b; León and Diez 1999; Sanches and Bellini 1999; Seminoff et al. 2003),
213 indicating varying periods of oceanic drifting prior to recruitment to neritic habitats
214 (Musick and Limpus 1997). For Cayman hawksbills, small size at initial capture
215 suggests that the oceanic phase may be relatively brief, and the continued capture of
216 small, unmarked juveniles throughout our study implies continual recruitment.
217 Indeed, significant growth in some Caribbean hawksbill breeding populations has
218 been observed (e.g. Beggs et al. 2007) – a trend which may ultimately be detected on
219 foraging aggregations.

220 In some areas, adult and juvenile hawksbill turtles share foraging grounds (Limpus
221 1992; Broderick et al. 1994; Musick and Limpus 1997), while in others, the larger
222 size classes are lacking (León and Diez 1999). The absence of the larger size classes
223 may represent an artifact of past exploitation (León and Diez 1999) but emerging
224 flipper tagging (Meylan 1999; Bellini et al. 2000; Grossman et al. 2007) and limited
225 satellite tracking results (Whiting and Koch 2006) suggest migrations from juvenile to
226 adult foraging grounds. In the Cayman Islands, the predominance of juveniles and a
227 tag return from Honduras suggests developmental migration of subadults – as is also
228 suspected for headstarted green turtles released from the Cayman Turtle Farm (Bell et
229 al. 2005).

230 Within the Cayman Islands, local recapture of marked hawksbills occurred over
231 periods of several years, indicating year round residence and long-term site fidelity in
232 some individuals. Compared to sites in Puerto Rico (Diez and van Dam 2002), body
233 condition index was relatively high but growth rate was slow. Because of slow growth
234 and extended residence in juvenile foraging habitat, anthropogenic and natural threats
235 will have a cumulative impact during the years that hawksbills are present in the
236 Cayman Islands – and stocks may have less resilience to exploitation.

Behavior

237 Turtle activity varied according to habitat: resting was frequently observed in
238 uncolonized areas and resting, swimming, and feeding were observed in coral reef,
239 reef wall, and colonized hardbottom habitats. In deeper waters, turtles were more
240 likely to be observed breathing, partly because when submerged they were likely out
241 of view, but perhaps also because they spent more time at the surface preparing for or
242 recovering from deep dives. Notably, little feeding was observed in the Cayman
243 Islands in comparison to Puerto Rico, where turtles fed almost continuously (van Dam
244 and Diez 1997). While turtles feeding by scraping the reef may be more easily
245 disturbed than turtles feeding on sponges, for the most part, feeding turtles did not
246 perceptibly react to the presence of observers or discontinue feeding.

247 Though hawksbills are generally considered solitary, aggressive interactions have
248 been documented in captivity, when two captured hawksbills were placed in a boat
249 together (Sanches and Bellini 1999) and in the wild, when two hawksbills attempted
250 to feed on the same sponge (van Dam and Diez 2000). In Little Cayman, two
251 aggressive – possibly territorial – interactions between hawksbills were observed
252 along the reef wall: in both incidences, hawksbills were observed biting a conspecific.

253 Images from underwater photographers offered an opportunity to qualitatively
254 enhance understanding of hawksbill habitat use and behavior. Turtles were
255 photographically documented at a variety of depths (including those beyond the range
256 of capture efforts). During the day, hawksbills were often observed in colonized
257 hardbottom, reef, and reef wall, and at night, they were seen wedged under ledges.
258 Thus, despite positive buoyancy, hawksbills may maximize dive duration and
259 minimize surfacing effort by “assisted resting” (Houghton et al. 2003) in shallow
260 water with fully-inflated lungs.

261 Observations of foraging behavior included feeding on sponges and scraping the
262 surface of the reef (a behavior which likely represents feeding on encrusting sponges,
263 invertebrates or algae; Carr and Stancyk 1975). Occasional consumption of thimble
264 jellyfish was also photographically documented. As jellyfish are digested more
265 rapidly than other food items, they may be under-represented in samples of stomach
266 contents. Thus, like deployment of animal-borne video cameras (Heithaus et al.
267 2002), photo-documentation can provide dietary insights. Additionally, a commensal
268 feeding relationship was noted between hawksbills and angelfish, in which angelfish
269 nibbled sponges where hawksbill feeding had exposed the interior tissues, or fed on
270 crumbs dropped by turtles.

271 Many Caribbean reefs, including those in the Cayman Islands, are heavily used for
272 dive tourism. While ecological impacts of recreational scuba diving (Tratalos and
273 Austin 2001) and potential disturbance of marine turtles by inwater activities
274 (Meadows 2004) are cause for concern, there is a corresponding but under-utilized
275 potential to harness recreational divers for biology and conservation (Bell et al. in
276 press). By requesting photographs from scuba divers, we were able to make use of the

277 immense number of hours they spend observing the behaviors of marine turtles in the
278 wild, without relying on anecdotal reports. Photographs collected from divers during
279 this project offered insights into hawksbill diet, habitat use, and behavior –
280 highlighting the utility of this technique in the study of charismatic marine animals.

Habitat use

281 For hawksbills in the Cayman Islands, distance traveled from capture to recapture was
282 comparable to other studies (van Dam and Diez 1998a; León and Diez 1999) and
283 individual turtles were recaptured in multiple habitats, suggesting that all of these
284 must be encompassed in the design of marine protected areas. There was no apparent
285 structuring of home range according to turtle size, but significant vertical structuring
286 was observed, with larger individuals generally captured in deeper waters. In Little
287 Cayman, hawksbill sightings and captures were clustered within narrow bands of
288 highly colonized habitat along the miniwall (former seacliff marking the transition
289 from shallow to deep terrace) and reef wall (near-vertical shelf edge). Given the
290 narrow shelf, diving down the face of the reef wall substantially increases available
291 habitat, and may buffer against anthropogenic and natural degradation of shallow
292 habitats (Blumenthal et al. 2009). Thus, as in other areas (Limpus 1992), vertical
293 features may provide critical habitat, yet these are necessarily under-represented on
294 two-dimensional habitat maps. Hawksbill density (catch per unit effort) was
295 significantly greater in Little Cayman. However, caution must be taken in comparing
296 catch per unit effort in this study with other studies, unless a similar method of towing
297 observers is used.

Caveats and considerations for future study

298 Our results illustrate how capture methods, survey design, and selection of study sites
299 may profoundly influence findings. Study sites differed significantly in physical
300 configuration: in Little Cayman, we were able to search the narrow shelf from near
301 shore to the shallow drop-off, while in Grand Cayman, the drop-off began in deeper
302 water. Likely due to lack of search effort in highly suitable deep habitats near the
303 shelf edge, captures in Grand Cayman were more widely dispersed and fewer large
304 turtles were captured. Generally, search effort in Little Cayman was more complete
305 and capture occasions were more efficient (i.e. catch per unit effort was higher). Also,
306 in Little Cayman, a much higher proportion of turtles were recaptured, allowing
307 estimates of growth and home range to be made. However, while Little Cayman
308 offered substantially greater insights into demographic parameters, it is possible that
309 these results are not representative of all areas. Therefore, a balance must be found
310 between monitoring a larger number of index sites – and diluting demographic data
311 by providing fewer opportunities to capture and recapture individuals.

312 In this study, effort was not uniform across habitat types, as we aimed to maximize
313 number of captures by searching more suitable areas. Additionally, a limited number
314 of captures were made per capture occasion, representing an efficient survey design
315 given resources and personnel, but precluding present estimation of population size
316 and survival by capture-mark-recapture modeling. In the future – and in designing
317 new monitoring programs – longer capture occasions could be undertaken, effort
318 could be expanded to include deeper waters using scuba methodology, and habitat
319 preference could be quantified using a random survey methodology. Based on our
320 calculations of size estimation accuracy, uncaptured turtles can be assigned to 10 cm
321 size classes – opening up the possibility of further studies of habitat use via sighting
322 transects.

Threats

323 Though historically the Cayman Islands were noted for abundant nesting by green,
324 loggerhead, leatherback and hawksbill turtles (Lewis 1940), migratory green and
325 loggerhead nesting populations are now critically reduced, and leatherback and
326 hawksbill nesting appears to have been extirpated (Aiken et al. 2001; Blumenthal et
327 al. 2006; Bell et al. 2007). Nevertheless, adult and subadult turtles, including
328 hawksbills, were captured in a traditional turtle fishery (Bell et al. 2006). In 2008, the
329 Cayman Islands government modified size limits for legal marine turtle take –
330 protecting vulnerable breeding populations but allowing smaller turtles to be targeted
331 for the first time in more than twenty years. However, a ban on legal take of hawksbill
332 turtles has been implemented, based on results of this study – showing slow growth
333 rate, long-term residence, and resultant vulnerability to anthropogenic threats. Thus,
334 by collecting diverse data on demographics, habitat use, behavior, and threats to
335 hawksbills in the Cayman Islands, we have informed local management and set a
336 baseline for an index inwater monitoring site in the western Caribbean.

Acknowledgements

337 We thank the following photographers for generously providing their images for
338 Figure 6: Michelle Foss, Gary Tayler, Patrick Weir <http://www.digitaldiver.biz>, Eric
339 Friberg, Gary Tayler, Katie and Chris Alpers <http://www.indigodivers.com>, Joanna
340 and Chris Humphries, and Alexander Mustard <http://www.amustard.com>. For
341 invaluable logistical support and assistance with fieldwork, we thank Department of
342 Environment research, administration, operations, and enforcement staff and
343 numerous volunteers. Work in the Cayman Islands and the UK was generously
344 supported by the National Fish and Wildlife Foundation (NFWF), the Turtles in the
345 Caribbean Overseas Territories (TCOT) project at University of Exeter, the UK
346 Department of Environment, Food and Rural Affairs (DEFRA), the National
347 Environment Research Council (NERC), the European Social Fund, and the Foreign
348 and Commonwealth Office for the Overseas Territories. We also acknowledge
349 support to Janice Blumenthal (University of Exeter postgraduate studentship and the
350 Darwin Initiative). The manuscript was improved by the comments from Robert van
351 Dam and anonymous reviewers.

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Figure legends

Fig. 1. Location of the three Cayman Islands in the western Caribbean sea. Inset shows the study areas of western Grand Cayman (GC) and Bloody Bay Little Cayman (LC). One hawksbill turtle tagged in LC was later captured in La Mosquitia, Honduras, suggesting a developmental habitat shift.

Fig. 2. Straight carapace length for hawksbill turtles captured (filled bars) and recaptured (unfilled bars) in a) Little Cayman and b) Grand Cayman.

Fig. 3. Straight-line displacement between capture and recapture points for hawksbill turtles. There was no significant correlation between displacement and turtle size (straight carapace length) or between displacement and time at large. Data for Little Cayman and Grand Cayman were combined.

Fig. 4 a) Turtle size (straight carapace length) and carapace length growth rate from first capture to last capture (cm yr^{-1}) were significantly correlated b) A highly significant correlation was observed between body mass and straight carapace length. Data for Little Cayman and Grand Cayman were combined.

Fig. 5. Activity with respect to habitat (uncolonized hardbottom, colonized hardbottom, reef, and reef wall). Turtles were observed feeding in colonized hardbottom, reef, and wall, while resting was more commonly observed in uncolonized habitats. Data for Little Cayman and Grand Cayman were combined.

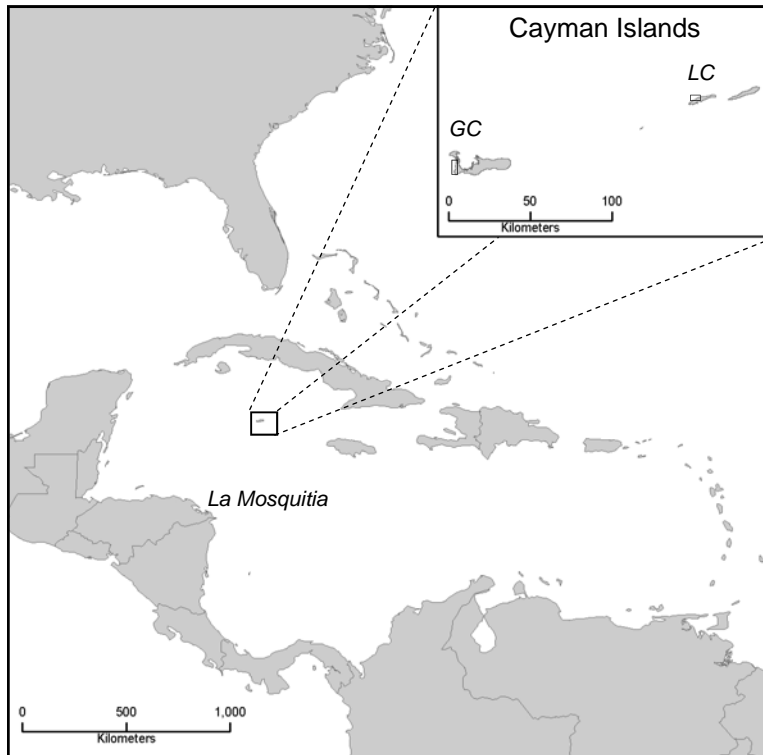
Fig. 6. Habitat use, feeding behaviors and interspecies feeding associations of hawksbill turtles in the Cayman Islands a) colonized hardbottom b) coral reef at the edge of the miniwall c) reef wall d) nocturnal resting wedged under a ledge e) hawksbill turtle scraping reef face f) angelfish feeding on interior sponge tissue

exposed a hawksbill g) angelfish feeding on crumbs dropped by a hawksbill h) hawksbill feeding on thimble jellyfish. Photographers: a) Michelle Foss b) Gary Tayler c) Patrick Weir d) Eric Friberg e) Gary Tayler f) Katie and Chris Alpers g) Joanna and Chris Humphries h) Alexander Mustard.

Fig. 7 a) Sighting locations for hawksbill turtles in Little Cayman, scaled according to turtle size (estimated straight carapace length), show lack of horizontal habitat structuring b) The significant relationship between turtle size (straight carapace length) and depth of water at the capture site indicates size-related vertical structuring.

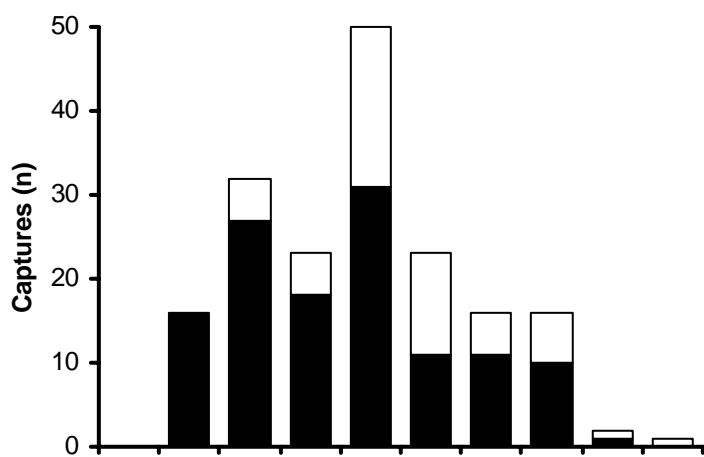
Fig. 8. Sighting and capture locations for hawksbill turtles with respect to habitats within the study areas of a) western Grand Cayman and b) Bloody Bay, Little Cayman. Captures in LC were clustered along vertical features (miniwall and reef wall) while captures in GC were more widely dispersed. Insets show the locations and geographic limits of the study areas.

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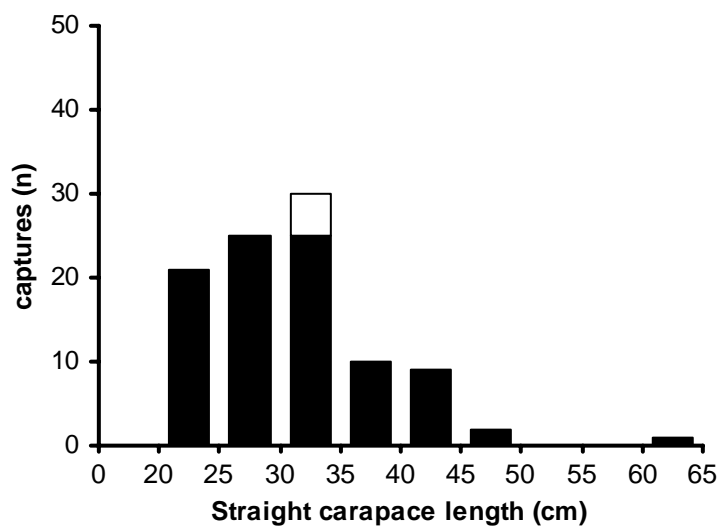


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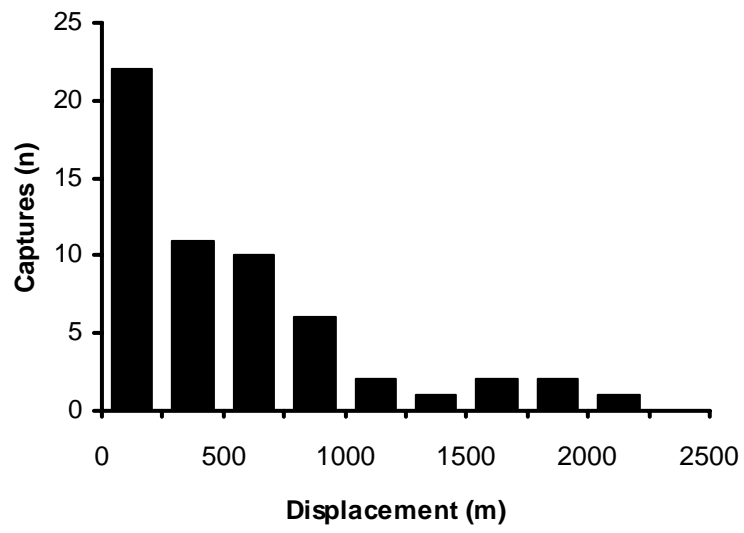
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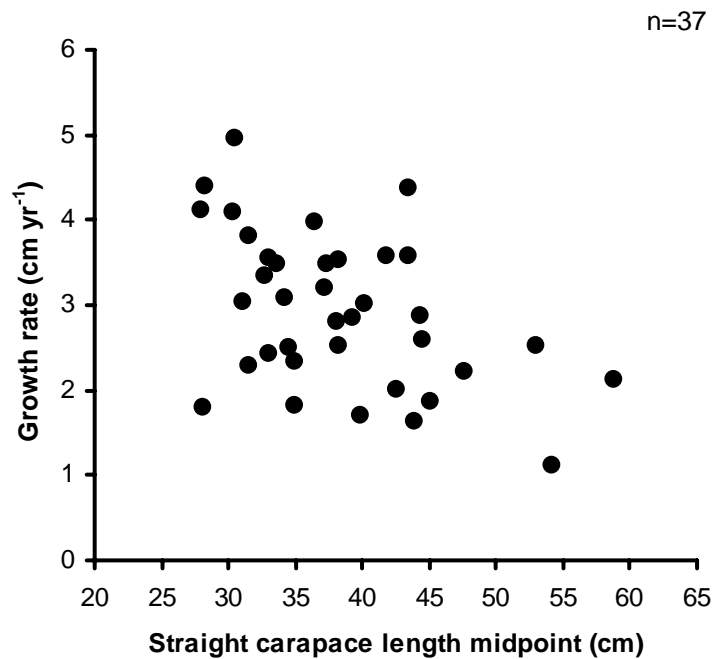


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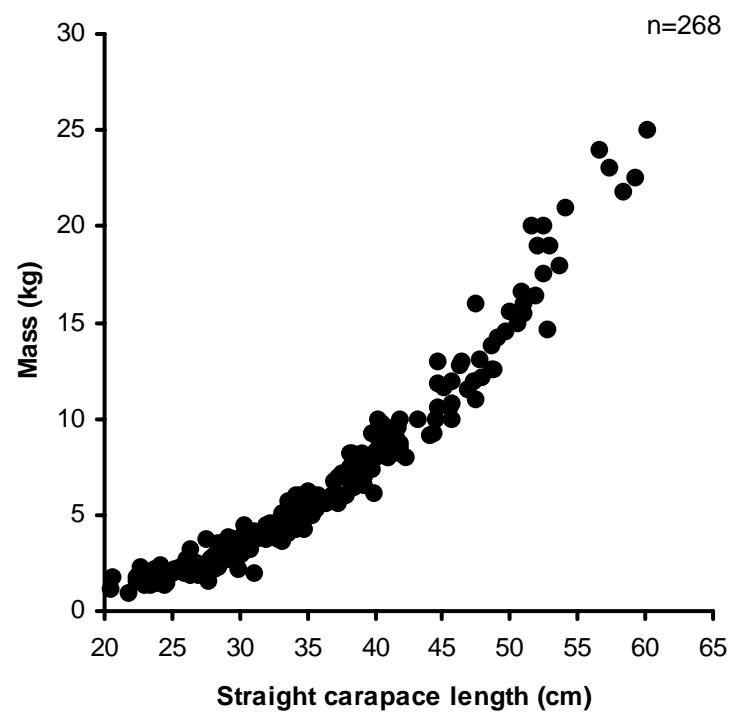


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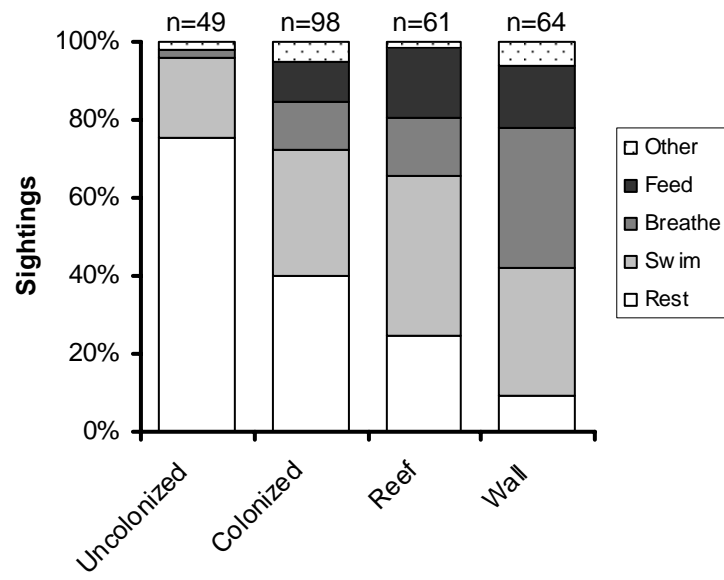
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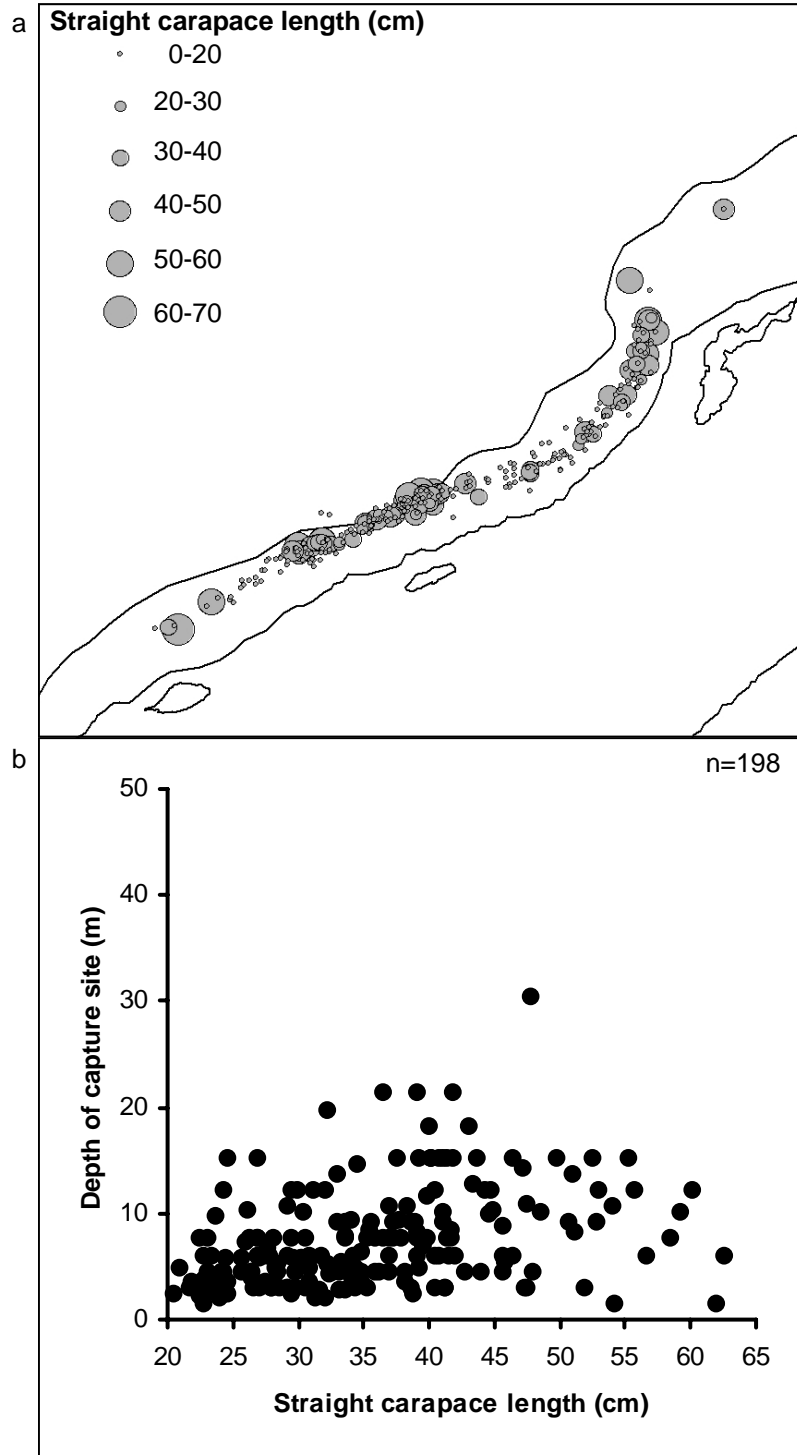
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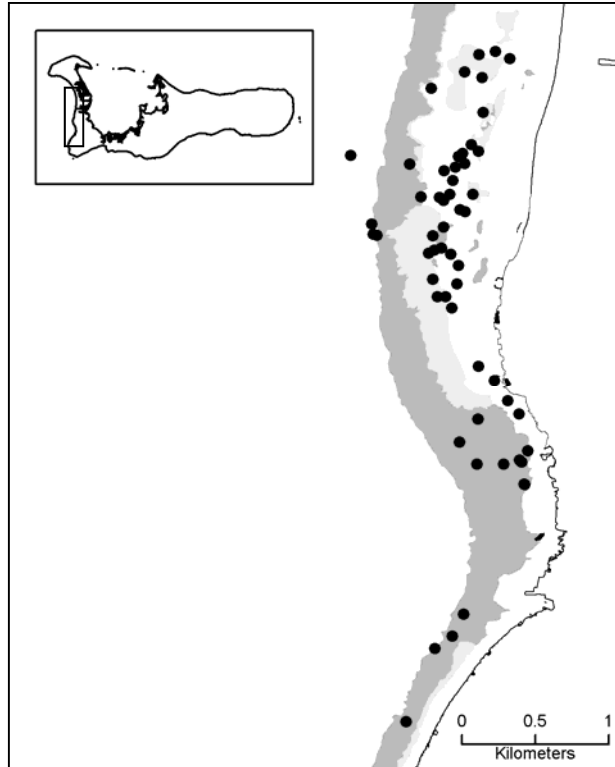


7.



8.

a



b

